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Small Mammal Habitat Use within Restored Riparian Habitats Adjacent to Channelized Streams in Mississippi

Peter C. Smiley Jr.[1*], Charles M. Cooper[2]

[1]USDA-ARS Soil Drainage Research Unit, Columbus, Ohio, USA; [2]USDA-ARS National Sedimentation Laboratory, Oxford, Mississippi, USA.
Email: *rocky.smiley@ars.usda.gov

ABSTRACT

Riparian zones of channelized agricultural streams in northwestern Mississippi typically consist of narrow vegetative corridors low in habitat diversity and lacking riparian wetlands. Land clearing practices and stream channelization have led to the development of gully erosion and further fragmentation of these degraded riparian zones. Currently, installation of a gully erosion control structure (drop pipe) at the riparian zone-agricultural field interface leads to the incidental establishment of four riparian habitat types that differ in habitat area, vegetative structure, and pool size. Small mammals were sampled within four sites of each habitat type from June 1994 to July 1995. Small mammal diversity, abundance, and hispid cotton rat (*Sigmodon hispidus*) weight were the least within smallest Type I habitats with the least vegetative structural diversity and were the greatest within the larger Type II, III, or IV habitats having greater vegetative structural diversity and pool size. Small mammal diversity and abundance were the least in the summer 1994, increased in the fall 1994, and then declined later in our study. Hispid cotton rat abundance was the least in summer 1994, winter 1994, and spring 1995 and was the greatest in fall 1994 and summer 1995. Our results suggest that modifying the drop pipe installation design to facilitate the development of larger riparian habitats with greater vegetative structural diversity will provide the greatest benefits for small mammals.

Keywords: Gully Erosion; Channelization; Channel Incision; Erosion Control; Small Mammals

1. Introduction

Within the last 200 years, >80% of riparian zones in North America and Europe have been destroyed and human induced modification of the remaining habitat continues [1]. Intact riparian zones are critical landscape features required for the maintenance of biodiversity [2]. Riparian zones exhibit high levels of plant and animal diversity, and may facilitate ecological and genetic exchange by serving as connectivity areas with other ecosystems [3]. When upland forest or grassland habitat bordering streams is cleared for agriculture, the remaining riparian zones may increase in significance as habitat for small mammals [4]. Riparian zones that differ in vegetative structure and species composition from adjacent upland habitats have greater diversity and abundance of small mammals [5-7].

Riparian zones adjacent to channelized agricultural streams in northwestern Mississippi consist of narrow vegetative corridors low in habitat diversity and lacking riparian wetlands. Agriculture and stream channelization have contributed to the degradation of these riparian zones since the 1830s [8]. The initial reduction of riparian zone width occurred as land adjacent to these streams was cleared and developed for agriculture. Federal channelization projects conducted between 1930 and 1960 initiated severe channel incision that resulted in the destabilization of entire watersheds [9]. Channel incision severs the typical floodplain-stream linkage and frequently results in gully erosion that rapidly migrates perpendicular to the stream through the riparian zone and into the agricultural field. Agricultural land in northwest Mississippi has been the location of some of the most active gully systems in the United States [10,11]. Gully erosion is the most severe form of soil erosion and can result in soil loss rates between 0.1 to 65 t·ha^{-1}·year^{-1} [12]. Fur-

*Corresponding author.

thermore, the creation of gaps caused by the loss of soil and riparian vegetation as a result of gully erosion leads to fragmentation of the riparian zones. Riparian habitat degradation resulting from channelization typically decreases small mammal diversity and abundance [13-15] and alters species composition [16]. The high frequency of gully erosion adjacent to channelized agricultural streams and the severity of the resulting erosion have resulted in the installation of thousands of drop pipes in northwest Mississippi [9,17].

Installation of drop pipes is often supported on a cost-share basis by the USDA Natural Resources Conservation Service, US Army Corps of Engineers, and other action agencies in the United States [18]. The drop pipe structure consists of an earthen dam placed across eroding gullies with an embedded L-shaped metal pipe that transports runoff to the stream (**Figure 1**). Drop pipe installation halts gully erosion and allows for the incidental development of riparian habitat that reconnects riparian zones fragmented by gully erosion [19,20]. Previous assessments have documented differences in fish, amphibian, reptile, bird communities, and avian nest predators among riparian habitats established by drop pipes [17,21-24]. Previous findings also confirmed that forested riparian wetlands had greater vertebrate species richness (*i.e.*, combined species richness of all vertebrate classes) than riparian habitats impacted by gully erosion and terrestrial riparian habitats created by drop pipe installation [19,25]. Differences in small mammal communities and populations among riparian habitats established by drop pipe installation have not been examined.

Previous studies [15,26] have evaluated small mammal responses to streamside management zones and stream restoration within forested watersheds in the southeastern United States. However, information on small mammal which responses to riparian habitat restoration within

Figure 1. Aerial photograph of newly installed drop pipe structure.

agricultural watersheds in this region is lacking. We sampled small mammals over a one-year period from four riparian habitat types established as a result of drop pipe installation in northwestern Mississippi to examine small mammal habitat use within these restored riparian habitats. Specifically, our research questions were: 1) Does small mammal community structure differ among habitat type and season?; and 2) Does the abundance, weight, number of recaptures, and movement patterns of hispid cotton rats (*Sigmodon hispidus*) differ among habitat type and season?

2. Materials and Methods

2.1. Habitat Classification and Description

A pre-study survey of 180 drop pipe sites within the Yazoo River watershed indicated that restored riparian habitats fit one of four discrete types on the basis of habitat area, pool volume, and vegetative structure (**Table 1**). Subsequent plant censuses and total station surveys [17] supported our initial habitat classification. Type I habitats were the smallest riparian patches and were composed mostly of herbaceous vegetation (**Table 1**). The four dominant plant species within Type I habitats were bermuda grass (*Cynodon dactylon*), goldenrod (*Solidago* spp.), paspalum grass (*Paspalum* spp.), and panic grass (*Panicum* spp.). Type II habitats were larger riparian patches than Type I habitats (**Table 1**) and were composed of herbaceous vegetation mixed with shrubs and saplings. The four dominant plant species within Type II habitats were Japanese honeysuckle (*Lonicera japonica*), goldenrod, white ash (*Fraxinus americana*), and blackberry (*Rubus argutus*). Type III habitats were riparian patches larger than Type II habitats (**Table 1**) and characterized by the presence of an ephemeral pool surrounded by a ring of woody vegetation. The four most frequently occurring plant species were black willow (*Salix nigra*), bermuda grass, ragweed (*Ambrosia artemisiifolia*), and kudzu (*Pueraria lobata*). Type IV habitats were characterized by having the greatest habitat area, permanent pools, greatest plant species richness (**Table 1**), and an input channel extending into the adjacent field. Vegetation within Type IV habitats consisted of woody and herbaceous vegetation, and the four most frequently occurring plant species were blackberry, goldenrod, partridge pea (*Cassia fasiculata*), and bermuda grass. The amount of woody vegetation within a site varied more among Type IV habitats than Type III habitats. Type IV habitats ranged from sites composed of predominantly herbaceous vegetation with a few mature trees > 2 m tall to sites that contained pools and input channels surrounded by mature trees > 2 m tall. Type IV habitats contained the largest trees and common woody species included black willow, American elm (*Ulmus*

Table 1. Mean total habitat area (m^2), mean maximum pool volume (m^3), mean plant species richness (plant richness), and mean vertical structure of woody vegetation index within four restored riparian habitat types in northwestern Mississippi, June 1994 to July 1995.

Habitat Type	Habitat Area	Pool Volume	Plant Richness	Vertical Structure Index[b]
Type I (n = 4)	600	15	22	0.01
Type II (n = 4)	1000	41	22	0.21
Type III (n = 4)	1300	426	26	0.30
Type IV (n = 4)	3700	1343	46	0.20

[a]See [17] for description of sampling methods for all response variables; [b]Index of vertical structure indicates dominance of woody vegetation greater than 1.8 m tall and ranges in scores from 0 (site lacking woody vegetation > 1.8 m tall) to 1 (site dominated by woody vegetation > 1.8 m tall) [17].

americana), American sycamore (*Platanus occidentalis*), and sweet gum (*Liquidambar styraciflua*).

The pre-study survey of 180 drop pipe sites also found that Type I habitats occurred most frequently (61%), followed in abundance by Type III (21%), Type II (11%), and Type IV (7%) habitats [9]. We selected four sites of each habitat type within the Long and Hotophia Creek watersheds in Panola County, Mississippi (lat 34°9' to 34°33N, long 89°43' to 90°11'W) as study sites (total 16 sites). Both watersheds were predominantly agricultural watersheds primarily devoted to cotton (*Gossypium hirsutum*) production. All study sites were adjacent to channelized agricultural streams and agricultural fields. We attempted to control for the potential influence of land use by choosing sites that were adjacent to cotton fields. Fifteen sites were adjacent to cotton fields, but logistical matters required us to select one site adjacent to a corn (*Zea mays*) field.

2.2. Small Mammal Trapping

Small mammals were sampled from 28 June 1994 to 20 July 1995 with sherman folding live traps (7.6 cm wide by 8.9 cm tall by 22.9 cm long) and pitfall traps (19 L buckets that are 34.9 cm tall and 28.6 cm in diameter) buried flush with the ground. The combined use of live traps and pitfall traps enabled us to sample the entire small mammal community more effectively. Pools within Type III and Type IV habitats prevented the use of the standard grid trapping array. Therefore, we established a trapping transect along the perimeter of each Type I, II, III, and IV site. The trapping transect encircled the entire site of Type I, II, and III sites. The large size of Type IV sites made surrounding the entire site impractical and the trapping transect was placed so it encircled the large pools within these sites. Placement of live traps and pitfall traps were alternated along the trapping transect and each trap was located 5 m apart. All traps were at least 1 m from the edge of all habitat types and the adjacent agricultural field. The total number of traps within a site varied, but the number of traps per transect length required to encircle the site (Type I, II, III) or pool (Type IV) was similar. Wood covers were placed above pitfall

traps and aluminum covers were placed over live traps to reduce trap mortality. Pitfall traps were unbaited and live traps were baited with a combination of peanut butter, rolled oats, and mixed seed the first trapping period of summer 1994. Rolled oats were used as bait for subsequent trapping periods because of problems with ants.

Trapping periods were summer 1994 (28 June to 1 July 1994; 18 July to 27 July 1994), fall 1994 (8 November to 16 November 1994), winter 1995 (21 February to 28 February 1995), spring 1995 (17 April to 26 April 1995), and summer 1995 (13 July to 20 July 1995). Trapping was conducted for 7 to 9 nights during each season within all sites. Traps were checked daily during each season and were closed between seasons. All animals captured were identified to species and released. Cotton rats were also weighed with spring scales and individually marked with ear tags (#1, National Band and Tag Co.).

2.3. Data Analyses

We calculated species richness (the number of species captured), evenness (the reciprocal of the Simpson's Index divided by species richness) [27], Shannon Diversity Index [28], and abundance (the number of captures) from every site during each season. Our use of two sampling methods with different capture efficiencies and the variability in sampling effort among habitat types and seasons (**Table 2**) required the use of an index to standardize our sampling effort. First, the sampling effort (*i.e.*, number of live trap nights and number of pit fall trap nights) within all sites and sampling periods for each sampling method was ranked. The sampling effort index is then calculated by summing the rank sampling effort of all sampling methods used in each site during each season [24]. We then standardized species richness and abundance values by dividing the values of these variables by the sampling effort index.

We calculated cotton rat abundance (# of individuals captured/sampling effort index), percent of juveniles (# of individuals < 100 gm in weight/total # of individuals) captured, mean cotton rat weight, and mean number of

Table 2. Mean number of trap nights for live traps (LT) and pitfall traps (PF) within four restored riparian habitat types in northwestern Mississippi during each season from June 1994 to July 1995.

	Type I		Type II		Type III		Type IV	
	LT	PF	LT	PF	LT	PF	LT	PF
Summer 1994	22	24	66	68	96	94	105	110
Fall 1994	22	24	82	84	102	100	120	126
Winter 1995	19	21	72	74	89	88	105	110
Spring 1995	25	27	92	95	115	113	135	142
Summer 1995	19	21	72	74	89	88	105	110

recaptures from each site during each season. We also calculated the average distance moved by an individual cotton rat (AVDI) with at least one recapture within each site. AVDI is the sum of all distances moved by an individual cotton rat divided by the number of days elapsed between the first and last recapture within a season [29]. We chose an index that averages movements over time rather than one that averages movements over number of recaptures because all individuals may not be captured during every trapping period [29]. We then calculated the mean AVDI from each site during each season.

We used a two factor analysis of variance (ANOVA) coupled with the Student-Neuman-Keuls (SNK) test to examine if differences in community (species richness, evenness, Shannon Diversity Index, abundance) and population response variables (cotton rat abundance, weight, percent juveniles, number of recaptures, movement) occurred among habitat types and seasons. The assumptions of normality and/or equal variance were not met for any community response variable and four of the population response variables (cotton rat abundance, weight, percent juveniles, number of recaptures). We rank transformed these eight response variables to conduct the two factor ANOVA. The use of rank transformation in conjunction with a parametric test is the equivalent of a nonparametric two factor ANOVA [30]. Additionally, some sites did not yield any cotton rat recaptures and resulted in missing values that prevented us from testing if an interaction effect occurred within the ANOVA analyses of mean number of recaptures and mean AVDI. ANOVA analyses were conducted using SigmaStat 3.1 for Windows [31] and a significance level of $P < 0.05$. Detrended correspondence analyses (DCA) were conducted on percentages of captures of small mammal species from each season to examine if species composition differed among habitat types in each season. To reduce the influence of rare species on the DCA results we omitted species that occurred in < 10% of all collections and selected the option in PC-ORD that results in downweighting of species that occurred < the frequency of the most common species/5 (*i.e.*, <15 to 20%) [32]. DCA analyses were conducted using PC-ORD for Windows [32].

3. Results

3.1. Community Responses

Ten small mammal species from 1743 captures occurred within a total of 12,446 trap nights. The three most frequently captured small mammals in all habitat types were hispid cotton rat, woodland vole (*Microtus pinetorum*), and the marsh rice rat (*Oryzomys palustris*) (**Table 3**). Species richness, evenness, Shannon Diversity Index, and abundance differed ($P < 0.05$) among habitat types (**Table 4**). Mean species richness was the greatest in Type II habitats and the least in Type IV habitats (**Table 5**). Mean evenness and abundance were the greatest in Type III habitats and the least in Type I habitats (**Table 5**). Mean Shannon Diversity Index within Type II, III, and IV habitats was greater than diversity within Type I habitats (**Table 5**). Species richness, Shannon Diversity Index, and abundance also differed ($P < 0.05$) among seasons (**Table 4**). Mean species richness, Shannon Diversity Index, and abundance were the least in summer 1994, increased in the fall 1994, and then either declined in the spring 1995 (species richness), declined in the summer 1995 (Shannon Diversity Index), or did not decline during the remainder of the study (abundance) (**Table 5**). None of the community response variables exhibited a significant interaction effect of habitat type and season (**Table 4**). Species composition did not differ among habitat types due to the variability in small mammal species composition that occurred among sites within habitat types (**Figure 2**). Type I habitats exhibited greater within habitat variability in species composition than the other habitat types from summer 1994 to spring 1995 (**Figure 2**). Variability in species composition was reduced within all habitat types in the summer 1995 (**Figure 2**).

3.2. Population Responses

We captured 386 individual cotton rats within all habitat types, and we marked and released 200 cotton rats between summer 1994 and spring 1995. One hundred and twenty cotton rats were recaptured within the same site at least once during a season. Only 32 from a possible 184

Table 3. Number of small mammal species captured within four restored riparian habitat types in northwestern Mississippi, June 1994 to July 1995.

Species	Type I	Type II	Type III	Type IV
Hispid cotton rat (*Sigmodon hispidus*)	77	273	209	320
Woodland vole (*Microtus pinetorum*)	1	37	115	64
Marsh rice rat (*Oryzomys palustris*)	2	38	96	52
Cotton mouse (*Peromyscus gossypinus*)	1	50	60	29
White-footed mouse (*Peromyscus leucopus*)	2	29	39	32
House mouse (*Mus musculus*)	23	23	19	18
Southern short-tailed shrew (*Blarina carolinensis*)	6	20	34	22
Southeastern shrew (*Sorex longirostris*)	0	10	7	12
Least shrew (*Cryptotis parva*)	2	8	4	5
Golden mouse (*Ochrotomys nuttalli*)	0	0	3	1

Table 4. P values from two factor analysis of variance tests conducted to determine if small mammal community and population response variables differed among habitat type, season, or the interaction of habitat type and season. P values < 0.05 are bolded.

	Habitat Type	Season	Interaction Effect
Community			
Species richness	**0.049**	**<0.001**	**0.848**
Evenness	**0.035**	0.212	**0.684**
Shannon diversity index	**<0.001**	**<0.001**	**0.108**
Abundance	**0.040**	**<0.001**	**0.209**
Population (*Cotton Rat*)			
Abundance	0.071	**<0.001**	**0.175**
Percent juveniles	0.296	**<0.001**	0.032
Weight	**<0.001**	**<0.001**	<0.001
Number of recaptures	0.143	0.081	-
Average distance moved	0.080	0.916	-

Table 5. Habitat and season factor means (SD) for small mammal species richness, evenness, Shannon Diversity Index (H'), and abundance in restored riparian habitats in northwestern Mississippi, June 1994 to July 1995.

Factor	Level	Richness	Evenness	H'	Abundance
Habitat Type	Type I	0.07 (0.09) AB	0.26 (0.38) B	0.26 (0.42) B	0.28 (0.37) B
	Type II	0.06 (0.04) A	0.49 (0.28) AB	0.99 (0.53) A	0.30 (0.20) AB
	Type III	0.05 (0.03) AB	0.59 (0.23) A	1.09 (0.51) A	0.32 (0.18) A
	Type IV	0.04 (0.02) B	0.44 (0.29) AB	0.89 (0.62) A	0.23 (0.17) AB
Season	Summer 1994	0.02 (0.02) C	0.31 (0.43) A	0.30 (0.45) B	0.06 (0.06) B
	Fall 1994	0.07 (0.04) A	0.43 (0.27) A	1.04 (0.48) A	0.47 (0.24) A
	Winter 1994	0.08 (0.05) A	0.57 (0.20) A	1.19 (0.50) A	0.35 (0.20) A
	Spring 1995	0.05 (0.03) B	0.51 (0.29) A	1.06 (0.63) A	0.23 (0.13) A
	Summer 1995	0.06 (0.08) B	0.40 (0.31) A	0.44 (0.40) B	0.30 (0.30) A

[a]Different letters indicate differences (P < 0.05) in ranked means among habitat type or season for each response variable.

individuals (accounting for known trap mortalities occurring within a season) were recaptured within the same site at least once between seasons. None of the seasonal recaptures occurred within Type I habitats, 12 seasonal recaptures occurred within Type II habitats, nine occurred within Type III habitats, and 10 occurred within Type IV habitats. Cotton rat abundance did not differ (P > 0.05) among habitat types, but differed (P < 0.05) among seasons (**Table 4**). Mean cotton rat abundance was the least in summer 1994, winter 1994, and spring 1995 and was the greatest in fall 1994 and summer 1995 (**Figure 3**). Percent juvenile cotton rats and weight exhibited a significant (P < 0.05) interaction effect (**Table 4**) as trends in these response variables among habitat types
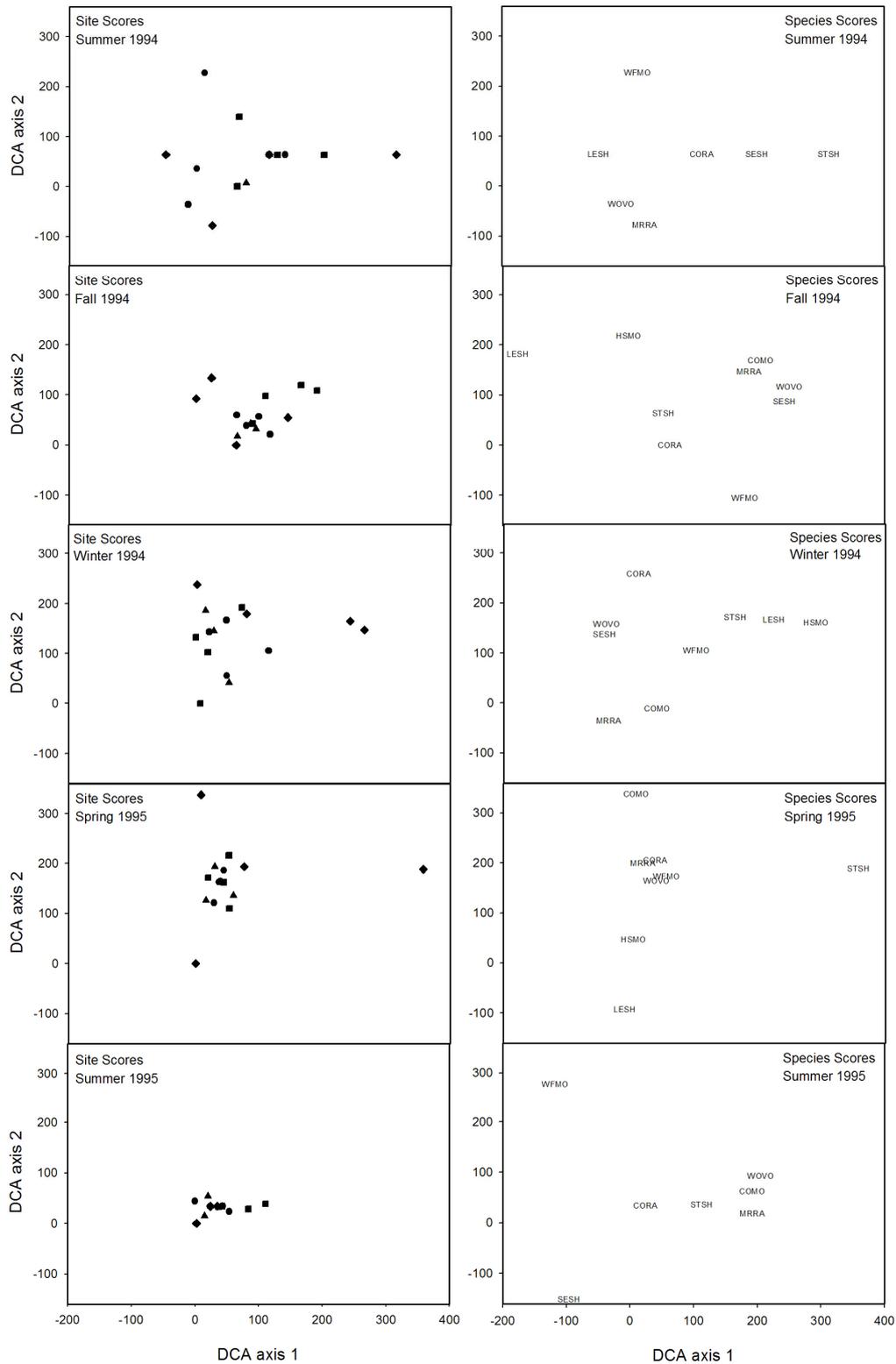
Figure 2. Site and species scores resulting from detrended correspondence analyses (DCA) of the percentage of small mammal species captured within four riparian habitat types in northwestern Mississippi during the summer 1994, fall 1994, winter 1994, spring 1995, and summer 1995. Habitat types are differentiated by different shapes within the figures: ♦: Type I sites; ●: Type II sites; ■: Type III sites; ▲: Type IV sites. Species codes are: COMO-cotton mouse; CORA-cotton rat; HSMO-house mouse; LESH-least shrew; MRRA-marsh rice rat; SESH-southeastern shrew; STSH-southern short-tailed shrew; WFMO-white-footed mouse; WOVO-woodland vole.
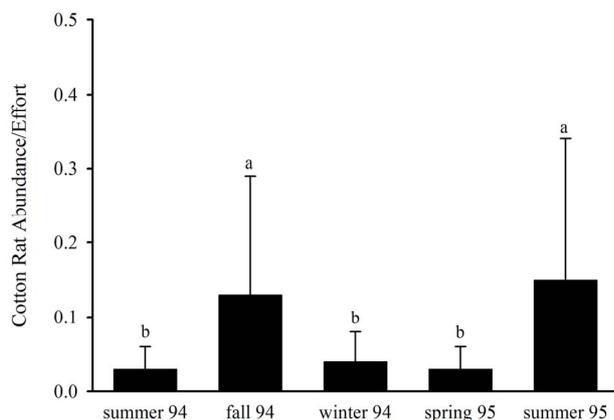
Figure 3. Mean cotton rat abundance within restored riparian habitats in northwestern Mississippi during summer 1994, fall 1994, winter 1994, spring 1995, and summer 1995.

differed among seasons. Mean percent juveniles did not differ among habitat types from summer 1994 to spring 1995 (**Figure 4**). Mean percent juveniles captured within Type I habitats was greater than the mean percent juveniles captured in the other three habitat types during summer 1995 (**Figure 4**). Mean cotton rat weight was greater in Type IV habitats than Type I, II, and III habitats during summer 1994 (**Figure 5**). No differences in mean cotton rat weight occurred among habitat types in fall 1994 (**Figure 5**). Mean weight was the least in Type I habitats, the greatest in Type II habitats, and intermediate in Type III and IV habitats in winter 1994 (**Figure 5**). Mean weight in Type II, III, and IV habitats was greater than mean weight in Type I habitats in spring and summer 1995 (**Figure 5**). Number of individuals recaptured and AVDI did not differ (P > 0.05) among habitats or seasons (**Table 3**).

4. Discussion

4.1. Community Responses

Riparian habitats established by drop pipe installation contained nine small mammal species of the possible 16 native species present within the watersheds of our study sites [33] and only one non-native species (house mouse) was captured. Small mammal diversity within the southeastern United States is depauperate compared to other regions of the United States [34]. The levels of small mammal diversity and species composition we documented within these restored riparian habitats was similar to those found in other created agricultural wetlands [35], restored riparian zones [34], and managed pine forests [26,35-39] in the southeastern United States. Native small mammal species not captured within our study sites included eastern mole (*Scalopus aquaticus*), eastern chipmunk (*Tamias striatus*), eastern gray squirrel (*Sciurus carolinensis*), fox squirrel (*Sciurus niger*), southern fly-

ing squirrel (*Glaucomys volans*), eastern harvest mouse (*Reithrodontomys humulis*), and the eastern woodrat (*Neotoma floridana*). Absence of these species from our study sites likely resulted from low capture vulnerability of these species to our trapping methods and/or their habitat preferences.

In general, Type II, III, and IV habitats exhibited greater species richness, evenness, diversity, and abundance than Type I habitats. We attribute these changes in community structure to the greater habitat area and vegetative structure within Type II, III, and IV habitats compared to Type I habitats. Small mammal communities within remnant habitats (*i.e.*, riparian zones, shelter belts, etc.) in agricultural landscapes are primarily influenced by habitat area and vegetative structure [40]. Small mammal species richness and abundance within riparian zones of agricultural streams in Quebec increased with increasing riparian widths and increasing vegetative structure [41]. Species richness of all small mammals and those preferring woodland habitat increased with increasing habitat area within fragmented forested riparian habitats and woodlots in Illinois [42]. Small mammal species richness and diversity increased with increasing habitat area and structural diversity within hedgerows in agricultural fields on Prince Edward Island, Canada [43] and within shelterbelts in Minnesota [40].

However, we did not observe a consecutive increase in species richness, evenness, diversity, and abundance from Type II to Type IV habitats despite increases in habitat area and vegetative structure. We were surprised that Type IV habitats with the greatest habitat area and vegetative structure did not exhibit the greatest species richness, diversity, and abundance values. We suspect these results stem from the influence of increasing pool development (*i.e.*, volume) that occurs from Type II to Type IV habitats. Little is known about the influence of increasing pool size on small mammal communities within riparian zones. Increased precipitation and flooding have been observed to decrease small mammal diversity and the abundance within riparian habitats in Mississippi and Missouri [44,45]. Increasing pool size within restored riparian habitats may also result in decreased small mammal diversity and abundance because it reduces the amount of terrestrial habitat within a site and the growth of herbaceous vegetation that is a critical habitat resource for many small mammals [45].

Seasonal changes in species richness, diversity, and species composition are likely a result of seasonal changes of cover and food resources within our sampling sites. We feel the seasonal changes in species richness and diversity were also influenced by the seasonal changes in cover and food resources within the adjacent agricultural fields. During the summer, crops present on the fields were approximately a meter tall, but little or no cover
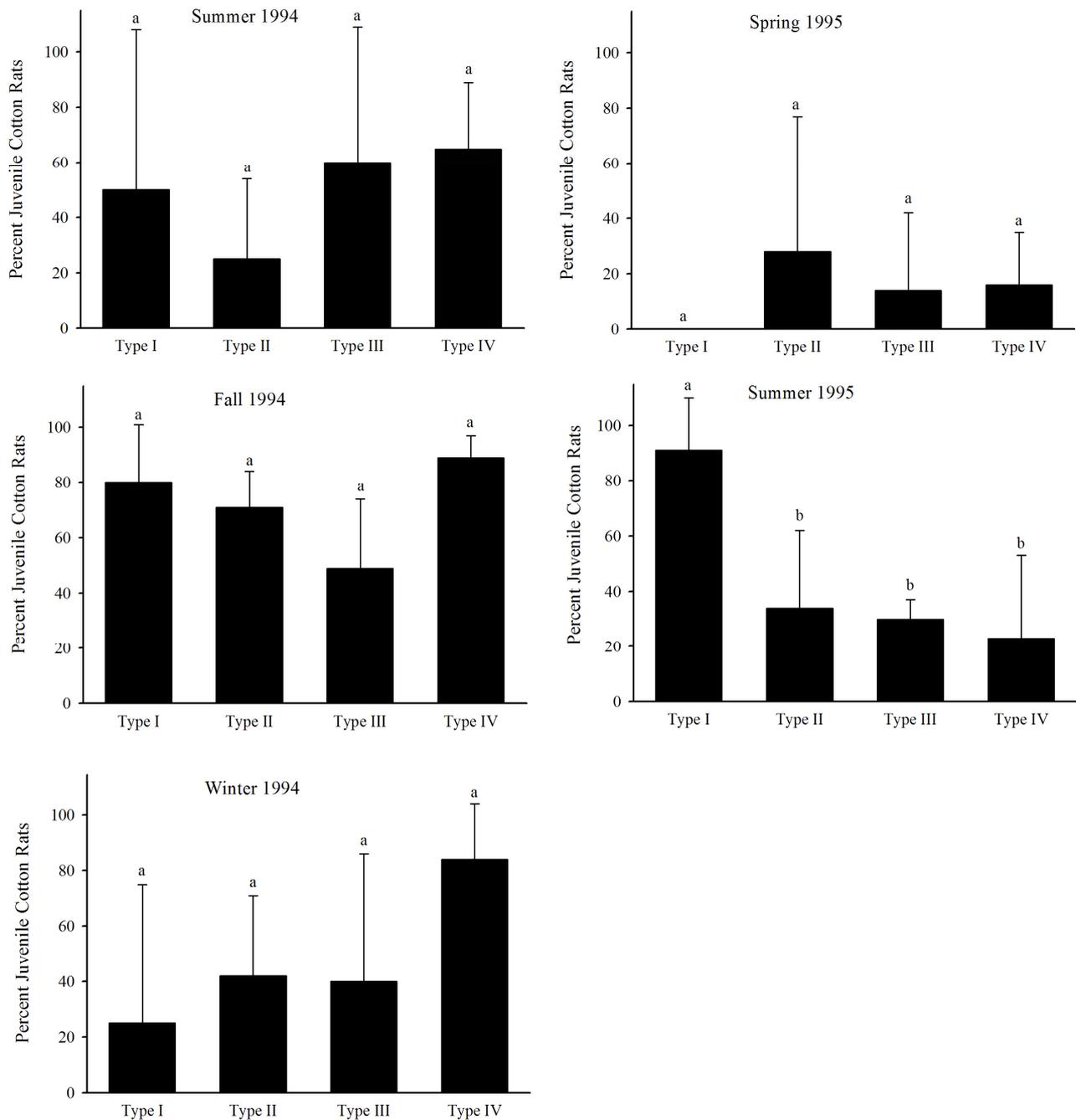
Figure 4. Mean percent juvenile cotton rats captured among four restored riparian habitat types in northwestern Mississippi during summer 1994, fall 1994, winter 1994, spring 1995, and summer 1995.

was present on the fields in the fall, winter, and spring. Therefore, observed declines in species richness and diversity during the summer may have occurred because small mammals were not limited to the riparian zone due to the increased cover and/or food resources available on agricultural fields.

4.2. Population Responses

Hispid cotton rats prefer grass dominated habitats, habitats with thick cover, and have been captured in vegetation areas adjacent to ponds [46,47]. Specifically, cotton rats prefer habitats having stands of dense grass species greater than 0.25 m high associated with various scrub species [46]. All habitats created by drop pipes met these general habitat criteria, which may explain the numerical dominance of cotton rats in all habitat types (**Table 3**). Cotton rat abundance, weight, and movement patterns were less within unpreferred habitats (mowed, minimal
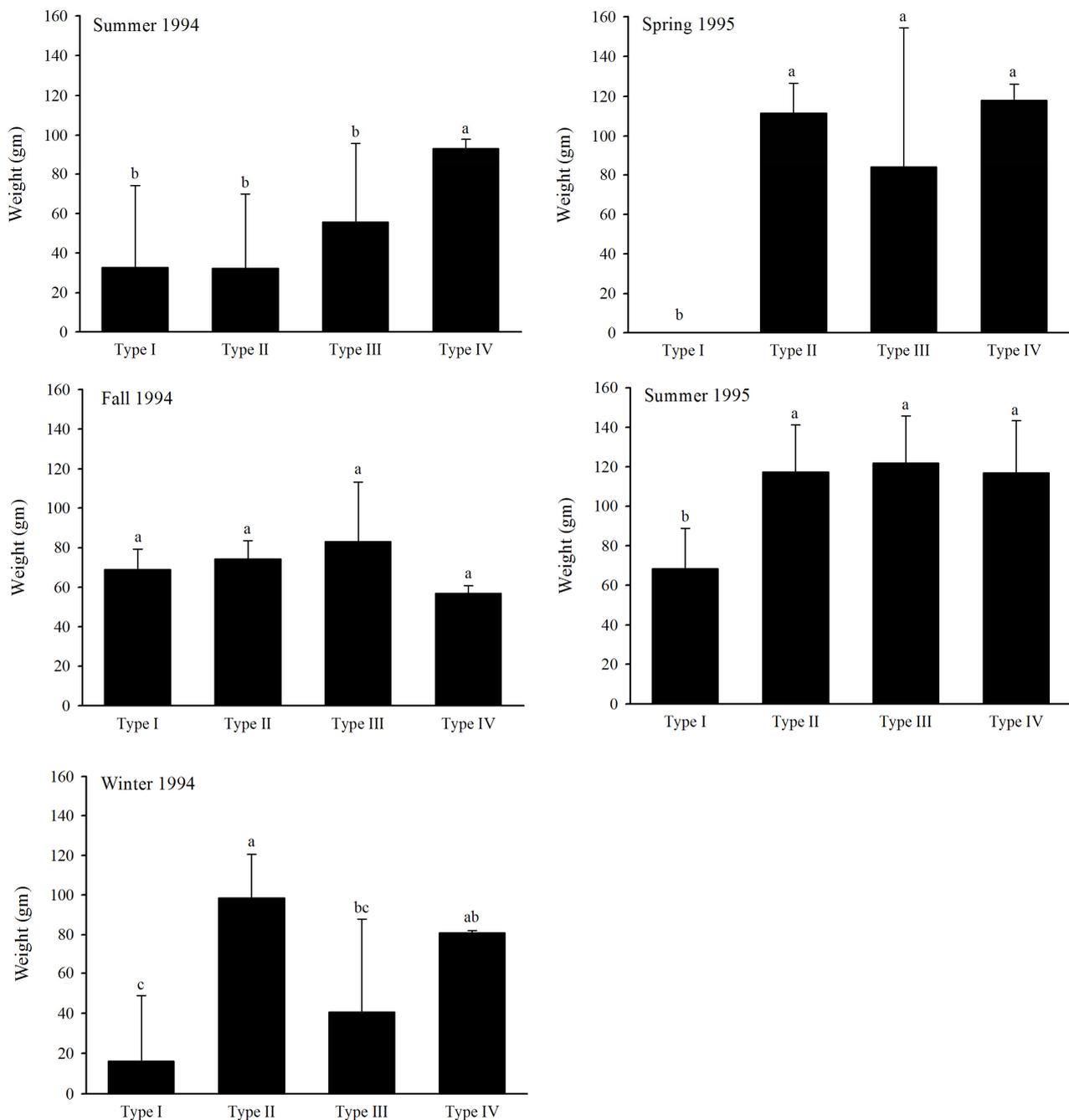
Figure 5. Mean cotton rat weight among four restored riparian habitat types in northwestern Mississippi during summer 1994, fall 1994, winter 1994, spring 1995, and summer 1995.

grass cover lacking shrub overstory) than preferred habitats (unmowed, dense grass cover with shrub overstory) in Texas coastal prairie [48]. Our results concur with these as we documented decreased cotton rat weights within small Type I habitats with reduced vegetative structure. We did not document differences in cotton rat abundance among restored riparian habitat types, but our results may have been obscured by gender differences as male and female cotton rats exhibit different habitat use patterns

[49,50]. We also documented greater percent of juvenile cotton rats in Type I habitats than the other three habitat types during summer 1995. Increased cotton rat abundance (adults and juveniles) also occurred during the summer 1995 and suggests the possibility that juvenile cotton rats were marginalized to the Type I habitats. Others in Oklahoma [46] and Georgia [49] have observed that smaller or juvenile cotton rats increase their usage of habitats with less cover during periods of increased

cotton rat abundances.

5. Conclusion

Environmental problems associated with stream channelization and subsequent channel incision and gully erosion occur within agricultural watersheds throughout the world [51,52]. Our results provide insights on how a gully erosion control structure and similar water control structures used nationally and internationally can be used to assist with the restoration of riparian zones of channelized agricultural streams. We observed that small mammal diversity, abundance, and hispid cotton rat weight were the greatest in the riparian habitat types (Type II, III, IV) having greater habitat size, vegetative structural diversity, and pool size than the smallest riparian habitat types (Type I) with the least vegetative diversity. Present drop pipe installation practices focus on erosion control without consideration of riparian restoration. Our results indicate small mammals will benefit the most from the establishment of Type II, III, and IV habitats that are the habitat types least frequently created by drop pipe installation. Modifying the drop pipe installation design to ensure that riparian habitats larger than 1000 m^2 with pool volumes greater than 41 m^3 and greater than 20% coverage of woody vegetation greater than 1.8 m tall are established within riparian zones adjacent to channelized agricultural streams in northwestern Mississippi will assist with the conservation of small mammals within these degraded riparian zones. Our findings also suggest that increasing the amount of terrestrial habitat associated with Type IV habitats containing large pools may promote increased diversity and abundance of small mammals. Future research on small mammals within riparian zones, agricultural fields, and other habitat types present within agricultural watersheds is necessary for developing watershed management plans that will benefit both wildlife and agriculture.

6. Acknowledgements

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