

# Biophysical Evaluation of the Vadose Zone at a Landfill Site in the Niger Delta, Nigeria

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## Abstract

The vadose zone of a landfill site proposed as an integrated waste management facility was evaluated based on geohydrological, chemical and microbiological characteristics of the groundwater and underlying soil. These data were also used to assess the attenuation capacity of the zone by the use of microbial degradation test of some major constituents including fatty acids, organic nitrogen and chloride of the leachate for a 28-day period. The main soil type in vadose zone consisted of brownish clayey sand of low permeability. The depth to water table which is equal to the thickness of the vadose zone varied from 8 - 13 m. Groundwater flowed with a hydraulic gradient of approximately  $4.0 \times 10^{-3}$  and a pore velocity of  $1.6 \times 10^{-3}$  cm/sec. The results of the biodegradation tests showed that the major constituents of the leachate such as ammonia/organic nitrogen, phosphate and organic carbon were completely degraded within 28 days. The population of aerobic bacteria within the 6 m soil depth was sufficient to bring about over 0.05% organic carbon removal. The soil characteristics in the vadose zone are very favourable for the occurrence of natural attenuation. The potential natural attenuation capacity of the vadose zone is therefore classified as moderate to high.

## Keywords

Attenuation, Biodegradation, Biophysical, Landfill, Vadose Zone

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## 1. Introduction

Landfills are covered or uncovered (open dump) waste receiving pits. They fall into different categories depending on the planning, design, construction, operation and closure as follows:

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1) Sanitary Landfills (SLF) are engineered sites where environment problems are minimized by a daily programme of spreading the received wastes in layers followed by compacting and then covering them with soils.

2) Secured Landfills are like sanitary SLF but are constructed such that there is no hydraulic connection between the wastes, leachate and natural waters.

3) Modern Landfills are well engineered facilities that are properly located, designed, operated, monitored and closed according to well defined procedures. Monitoring continues even after closure. They are cleaned when necessary and are fenced to ensure compliance with government regulations to protect human health, environment and groundwater resources.

In Nigeria, open dumping of wastes on any available land space is a common method of solid waste disposal by urban dwellers despite the pollution associated with this method compared to the environment friendly waste management techniques of waste avoidance, source reduction, reuse and recycling. The open dumps have their inherent disadvantages but a more serious long term threat is the production of leachate from chemical and biological processes that break down wastes. When water comes in contact with wastes at a dump site, a foul smelling noxious liquid called leachate is produced which may contain toxic materials leached from the wastes. Once spilled in the ground, these contaminants migrate downward through the underlying strata under gravity and capillarity. This movement is enhanced by many processes including molecular diffusion, advection/convection, and mechanical dispersion. Physical, chemical and biological interactions in the subsurface by soil and leachate may ultimately pollute the environment. This implies that potential adverse, long term and irreversible impacts may occur in groundwater systems since the chemical composition of the leachate may exceed permissible limits. In landfill construction, site selection is important because the biophysical conditions of the subsoil (stratification and type) and geohydrologic parameters (aquifer type, depth to water table, flow directions) must be demonstrated to conform to local requirements for environmental protection. Despite advances in landfill construction technology, long term impact on the structural competence of the site by composite geologic, chemical and biological characteristics of both soil and water are always factored into landfill design, thus a thorough investigation of a proposed landfill site is often carried out before construction [1]-[4].

Soils play an important role in groundwater protection and in the transport and fate of the contaminants. The movement and attenuation of contaminants in the subsurface environment depend on several geological, chemical and biological factors including the site geology, thickness of the vadose zone, groundwater flow direction and velocity, aquifer hydraulic properties and quality and abstraction of surface and groundwater. The site in the Niger Delta area of Southern Nigeria where a sanitary landfill is proposed to be constructed is a wetland with potentials for severe environmental impacts. The mean annual rainfall exceeds 2000 mm. Groundwater tables are often very close to the surface. Thus landfills with potentials to cause pollution need to be planned, designed, constructed, operated and maintained according to best practice [5]-[10].

Three kinds of studies are important of proposed landfill sites. First, the geohydrological integrity of a site must be ascertained to ensure that it is stable and ground water friendly. Geotechnical and hydrogeological data required at landfill sites include, bulk density of the soils, discontinuities, hydrogeological characteristics of vadose and soil and water geochemistry [11] [12]. Secondly, biological and chemical studies of the vadose zone are important to assess the natural attenuation capacity of the soils in the event of leakages from the facility into surrounding soils. Finally, the transport and fate of contaminants at the site are always modelled to predict the pathways and receptors of the contaminants over a period of time if pollution occurs. This evaluation of the complex site biophysical factors that control the transformation of waste to the final state is critical to the safe construction of sanitary landfill to ensure that soil and groundwater is protected from contamination from leachate and surface runoff [13]. The important site parameters that significantly influence the occurrence of natural attenuation can be subdivided into geological (hydrogeological), chemical and biological characteristics. In a previous study, [14] used only biological characteristics of soils to assess natural attenuation in the landfill site proposed as an integrated waste management facility for use in the disposal of general refuse and pre-treated hazardous material. The present study evaluates the subsurface biophysical conditions of soils and groundwater in the vadose zone as composite factors in the assessment of the parameters favourable to the potential occurrence and natural attenuation capacity at the site by using geological and chemical characteristics.

## 2. Study Area Description

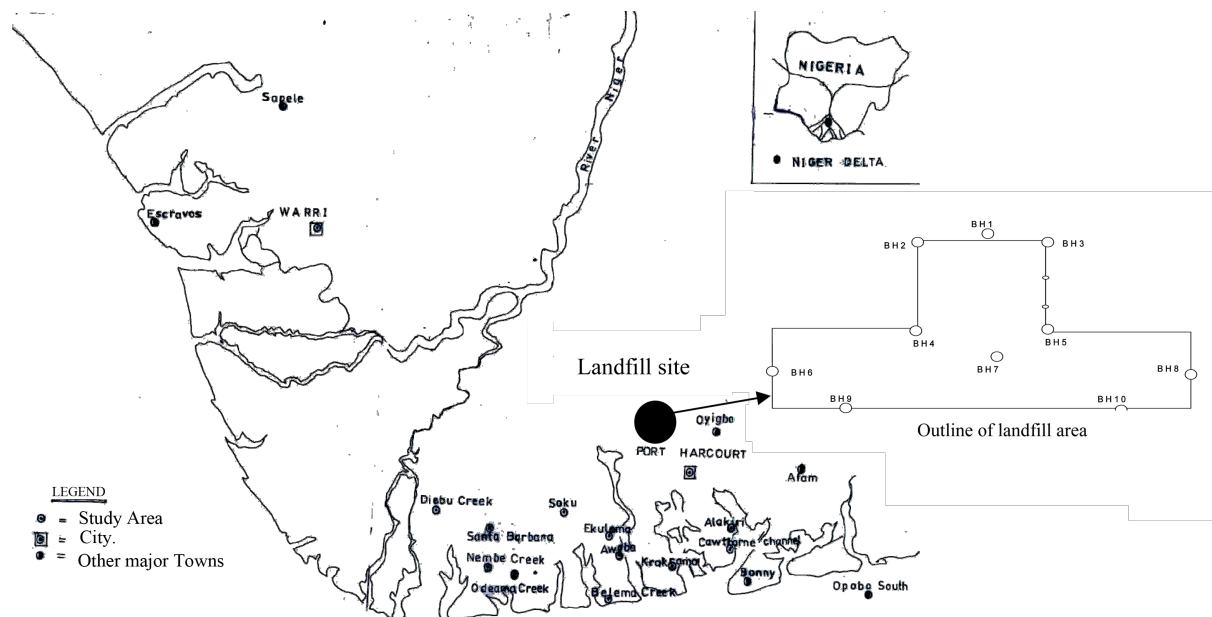
The landfill site is located 15 km north of Port Harcourt in southern Nigeria. Vegetation is characteristically rainforest with very tall trees. The area is drained by the Otamiri River which lies about 4 km to the west of the

proposed facility site which is also the local groundwater flow direction. Geologically, the site is part of the Tertiary Niger Delta. The local rock cover consists essentially of the Benin Formation which is predominantly sandy with a few shale intercalations. The sand and sandstones are coarse-grained and constitute the main regional aquifer. The location of the site and an outline of the landfill geometry are shown in **Figure 1**.

### 3. Materials and Methods

The study was undertaken towards the end of the wet season (October-November). Ten environmental boreholes, each 20 m deep, were drilled by the percussion method at strategic locations across the proposed integrated waste management facility as shown in **Figure 1**. Borehole 1 was specifically located slightly outside the study area at the up-gradient section of the project area to also serve as a control while boreholes 9 and 10 were located in the down-gradient to serve as monitoring observation wells. The wells were completed with 5 cm diameter PVC casing and screens. The annulus of the boreholes was backfilled with uniformly graded coarse sand up to 0.5 m above screen level. Well cuttings were used to backfill the remaining part of the annulus up to 0.75 m below ground surface and finally cement grouted to the surface, ensuring that the cement grout completely sealed the casing in order to prevent the introduction of surface contaminants into the aquifer. The wells were then developed and thereafter, water levels were allowed to stabilize for one week following which the static water levels were measured using a Fisher Model electric water level indicator before water samples were collected for analysis. To avoid contamination by drilling chemicals, soil samples for moisture content, physico-chemical properties and microbiological studies were collected at 0.5 m intervals from 3 m auger holes located beside each of the 10 boreholes. Undisturbed samples for bulk density determination were taken in a 50 m grid at the depth of 150 - 400 mm. In the laboratory, the geotechnical index properties of the soils including particle size distribution, bulk density, permeability, Atterberg limits, and porosity were determined as specified in [15]. Soil nutrients including pH, Organic matter,  $\text{PO}_4^{2+}$ ,  $\text{SO}_4^{2+}$ ,  $\text{Ca}^{2+}$ ,  $\text{Mg}^{2+}$ ,  $\text{NO}_3^-$ ,  $\text{NH}_4^+$ ,  $\text{NO}_2^-$  and  $\text{K}^+$  were determined using standard [16] methods, as well as heavy metals, total organic hydrocarbon and oil and grease. Total organic carbon and total organic nitrogen of soil were determined using Walkley-Black and Macro-Kjeldahl methods respectively. Soil pH was determined using PYE UNICAM PW 9418 pH meter fitted with a combined glass pH and reference electrode. Soil moisture content was determined by evaporation on Whatman filter paper No.1 at  $103^\circ\text{C}$  in an electrical oven.

The attenuation capacity of the vadose zone was evaluated using microbial degradation test of some major constituents of the leachate including fatty acids, organic nitrogen and chloride for a 28-day period. Olive oil served as source of fatty acid. Soil mixed with mercury chloride was used as negative control while surface soil



**Figure 1.** Map of the Niger Delta showing landfill site and its outline.

(0 - 15 cm depth) from the same area was used as control. Two concentrations of olive oil, ammonium sulphate and sodium chloride which served as sources of organic carbon, nitrogen and chloride respectively were used to simulate anticipated leachate from the facility. The objective of the experimental design was to simulate leachate concentrations of fatty acids, chloride and ammonium nitrogen that were similar to the anticipated leachate concentrations during the dry and wet seasons. Ten treatment options in triplicate were employed as follows:

Option 1, 2, 3, 4 had concentration of  $(\text{NH}_4)_2\text{SO}_4$  (olive oil) that were of dry season value.

Options 5, 6, 7, and 8 had concentration of olive oil that were of wet season value

Options 2, 4, 5, and 8 had concentration of  $(\text{NH}_4)_2\text{SO}_4$  that were of dry season value

Options 1, 3, 6, and 7 had concentrations of  $(\text{NH}_4)_2\text{SO}_4$  that were of wet season value

Options 3, 4, 7 and 8 had concentration of NaCl that were of dry season value while

Options 1, 2, 5, and 6 had concentration of NaCl that were of wet season value

Options 4 and 6 were representative of leachate during dry and wet season respectively.

Options 1, 2, 3, 5, 7 and 8 were mixed.

Option 9 was the negative control which contained olive oil and mercury chloride poison to kill bacteria.

Option 10 was the control sample which contained dry season levels of olive oil  $(\text{NH}_4)_2\text{SO}_4$  and NaCl in top soil (0 - 15 cm depth) which is expected to contain more bacteria than the deeper soil. The total heterotrophic bacteria (THB) count of soil samples was performed by inoculating 0.1ml of appropriate dilution of soil suspension in deionized water on nutrient agar plates using the spread plate technique [17].

Higher concentrations of constituents in the 10 options represented dry season levels, while the lower concentrations represented the wet season levels. Each option was placed in a dark plastic container of 1.5 L volume biodegradation test vessel and then kept in a thermos cooler which contained a lit candle to eliminate most of the air for the 28-day duration of the study. The candle light was extinguished after 3 hours as a result of depletion of oxygen level with the exception of option 10. The objective of this low oxygen incubation at room temperature was to simulate microaerophilic to anaerobic condition in the soil (at depth 6.5 m) where low levels of oxygen are characteristic. The options were watered every 7 days with 25 ml of distilled water to maintain a fairly constant moisture content of soil in all the options. Analysis of soil physicochemical properties and total heterotrophic bacteria count was performed on surface and 6.5 m deep soils before and after treatment with simulated leachate on day 0, 7, 14, 21, and 28. The physical and chemical characteristics of the soils before treatment are shown in **Table 1**. Qualitative evaluation of the attenuation capacity involved the overall assessment of the site

**Table 1.** Physical properties of soil before treatment.

Parameters	Soil samples	
	(0 - 15 cm depth)	(6.5 m depth)
Description	Darkish top soil	Brownish clayey soil
Porosity	0.38	0.38
Bulk density ( $\text{kg/m}^3$ )	1750	1890
Permeability (cm/sec)	$3.6 \times 10^{-5}$	0.0144
Sand	75%	75%
Silt	20%	20%
Clay	5%	5%
pH	$6.65 \pm 0.03$	$6.2 \pm 0.2$
% Moisture	$14.4 \pm 0.5$	$10.7 \pm 0.03$
% Total Organic Carbon	$0.58 \pm 0.04$	$0.18 \pm 0.02$
% Organic Nitrogen	$0.105 \pm 0.02$	$0.0035 \pm 0.0001$
% Ammonium Nitrogen	$0.02 \pm 0.01$	$0.0014 \pm 0.0001$
Chloride (mg/g)	$780 \pm 10.5$	$750 \pm 0.5$
Sulphate (mg/g)	$120 \pm 5.0$	$71.4 \pm 2.5$
Phosphate (mg/g)	$12.0 \pm 0.8$	$1.6 \pm 0.2$

conditions using hydrogeological, chemical and biological characteristics which are considered as important site parameters that influence the occurrence of natural attenuation.

#### 4. Results

The stratigraphy of the area delineated from soil samples obtained in the 20 m deep boreholes is characterized by three major soil types: a 1-m thick top darkish clayey fine sand underlain by a 10-m thick brownish clayey sand which is in turn succeeded by a 9-m thick fine to coarse sand horizon. The characteristics of the various horizons are presented in the **Table 2**. At the time of the investigation in October which is about the end of the rainy season, the depth to water table which is equal to the thickness of the vadose zone varied from 8 - 13 m. The groundwater flows in a westerly direction towards the Otamiri River with a hydraulic gradient of approximately  $4.0 \times 10^{-3}$  (or 0.4%). Since the average hydraulic conductivity of the aquifer is  $3.5 \times 10^{-3}$  cm/sec and the effective porosity is 0.45%, the pore velocity of the groundwater is therefore equal to  $1.6 \times 10^{-3}$  cm/sec. The [18] which regulates environmental guidelines and standards for the petroleum industry in Nigeria specifies that the top of a landfill should be at least 1.5 m below the land surface and the bottom should be 1.5 m above the water table. When these depths (1.5 m + 1.5 m) are added to the proposed 5 m thickness of the landfill, the minimum thickness of the vadose zone that is required is 8m which is therefore adequate at this location. The presence of clay minerals will decrease not only the permeability of the soil but also its absorptive capacity. The depth to the water table and the clayey nature of the soil implies that downward migration of contaminants may be hindered in the vadose zone. On contact with groundwater however, the pollution plume and dissolved phases will move in the direction of the groundwater flow by advection.

The groundwater chemistry is shown in **Table 3**. The groundwater is slightly basic in nature with pH values ranging from 6.97 to 7.25. Only water in BH 6 and BH 9 exhibited pH values that were below 7.0. The conductivity values range from 80.2  $\mu$ S/cm to 287.0  $\mu$ S/cm indicating low ionic content of the groundwater. Sulphate ions ranged from 0.02 - 0.62 mg/l. The concentrations of nitrate and phosphate are much lower than that of sulphate. Most of these parameters are within the [19] limits for potable water.

**Table 2.** Index properties of the various soil layers.

Layer	Depth (m)	Description	Texture/percentage composition	Uniformity coefficient	Permeability (cm/sec)	Liquid Limit (%)	Plastic Limit (%)	Plasticity index	Natural Moisture content (%)
I	0 - 0.5	clayey fine sand)	Sand (65% - 76%) Silt & clay (24% - 35%)	-	$10^{-5}$	34 - 45	17 - 21	17 - 24	14 - 35
II	0.5 - 11	Brownish clayey sand	Sand (75% - 96%) Silt & clay (4% - 25%)	4	$10^{-3} - 10^{-4}$	40 - 44	21 - 22	19 - 22	15 - 37
III	11 - 20	fine-coarse Sand	Sand (99% - 100%) Silt & clay (0% - 1%)	4 - 6	$10^{-2} - 10^{-3}$	33 - 40	15 - 18	16 - 22	12 - 22

**Table 3.** Physico-chemical characteristics of groundwater.

BH No	p <sup>H</sup>	Cond. ( $\mu$ S/cm)	TDS (mg/l)	DO mg/l	Temp °C	TSS mg/l	BOD (mg/l)	Turbidity (NTU)	Hardn (mg/l)	Cl <sup>-</sup> mg/l	SO <sub>4</sub> <sup>2-</sup> mg/l	NO <sub>3</sub> <sup>-</sup> mg/l	PO <sub>3</sub> <sup>4-</sup> mg/l
BH1	7.11	287	142	4.8	28	650	15	1.60	65	4.5	0.02	0.09	0.06
BH3	7.11	80.2	42.4	5.6	28	7.30	50	2.30	95	3.0	0.04	0.04	0.03
BH4	7.17	119.7	55.8	5.5	27.5	15.60	70	3.5	75	4.0	0.62	0.07	0.05
BH5	6.99	282	140	4.8	28	460	180	2.8	80	5.0	0.06	0.10	0.07
BH6	7.25	125.9	62.9	5.3	28.5	10.40	65	3.2	85	4.5	0.06	0.08	0.05
BH8	6.97	185.8	93.0	5.2	27.5	20.80	190	2.6	255	3.5	0.04	0.06	0.007
BH9	7.13	84.6	42.1	5.8	28.0	12.20	150	2.6	55	23.8	0.04	0.05	0.04
BH10	7.04	115.6	57.9	5.4	27.5	5.63	75	3.0	75	3.5	0.04	0.08	0.06

The mean microbial count in the soil is summarized in **Table 4**. The geological and chemical data were integrated with biological characteristics used by [14] to evaluate the natural attenuation potentials of the vadose.

## 5. Discussion of Biodegradation Results

The percentage reduction in the various properties with time during the degradation period is presented in **Figure 2** and the overall results summarized in **Table 5**. The reductions in the contents of the various components including ammonia nitrogen, organic nitrogen, chloride, sulphate and phosphate indicate that they were utilized during the biodegradation process. Relatively smaller reductions in the chloride and sulphate concentrations indicate they are less significant in the biodegradation process compared to ammonia nitrogen, organic nitrogen, and phosphate. pH declined progressively in all treatment options (except option 9) due to the microaerophilic and anaerobic degradation of olive oil into fatty acids. The fatty acid may have accentuated the lowering of the pH. Where the conditions were mainly aerobic and microaerophilic as obtained in option 10, the degradation of olive oil first gave rise to fatty acids before complete mineralization occurred in the presence of oxygen. In the dry season (option 4), the pH is lower because there were higher concentrations of olive oil,  $(\text{NH}_4)_2\text{SO}_4$  and NaCl compared to the wet season.

Ammonium nitrogen and organic nitrogen decreased steadily in all the options except option 9) due to microbial utilization of nitrogen [20] [21]. Organic nitrogen utilization was greater in the dry than wet season implying that the rate of organic nitrogen increased with high concentrations of utilizable nitrogen source. In the field, the availability of utilizable nitrogen source in a leachate is higher in the dry season than in the wet season when dilution occurs. The seasonal differences in the contents of the other components may similarly be explained by

**Table 4.** Mean microbial count of soil samples.

S/No	Sample Description	Mean Microbial count (cfu/g. or ml. of sample)			
		Heterotrophic count	Fungal count	Coliform count	Hydrocarbon degraders
1.	Surface (Soil, 0 - 15 cm)	$8.30 \times 10^5$	$1.44 \times 10^4$	$< \times 10^3$	$3.7 \times 10^4$
2.	Subsurface (Soil, 0.5 m)	$3.86 \times 10^4$	$2.13 \times 10^3$	$< \times 10^3$	$1.0 \times 10^3$
3.	Subsurface (Soil, 1.0 m)	$5.87 \times 10^3$	$< \times 10^3$	$< \times 10^3$	$1 \times 10^3$
4.	Subsurface (Soil, 1.5 - 3 m)	$1 \times 10^3$	$< \times 10^3$	$< \times 10^3$	$1 \times 10^3$
5.	Subsurface (aquifer mud)	$4.80 \times 10^5$	$8.53 \times 10^3$	$< \times 10^3$	$1.08 \times 10^4$
6.	Subsurface (aquifer water)	$4.78 \times 10^3$	$\times 10^3$	$< \times 10^3$	$1 \times 10^3$

**Table 5.** Percent reduction in soil parameters during degradation period.

	Dry season samples				Wet season samples			
	% reduction from weeks 1 to 4				% reduction from weeks 1 to 4			
Week	1	2	3	4	1	2	3	4
PH	3	8	8	8	2	2	3	4
Soil moisture	10		10		10	10	10	
Ammonia Nitrogen	29	63	82	100	26	66	84	100
Organic nitrogen	26	48	75	97	17	26	30	41
Chloride	5	5	5	52	0	0	0	20
Phosphate	57	87	92	95	30	55	80	100
Sulphate	16	35	44	50	3	7	6	7
Organic carbon	49	94	95	95	50	99	100	100
THB (increase)	900	2100	20,900	79,900	1500	3900	15,900	74,900

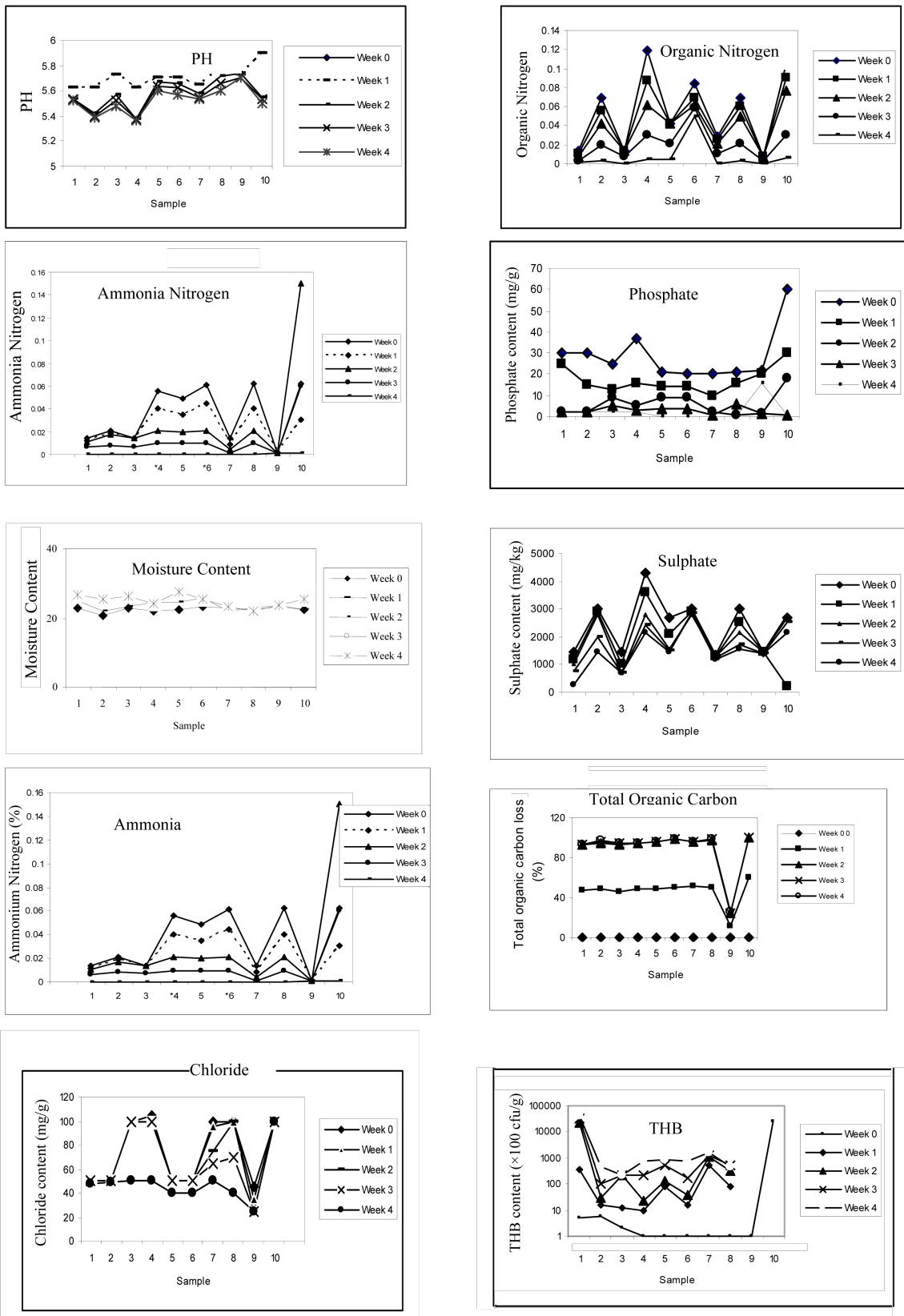


Figure 2. Changes in the components of the soil during degradation period.

this phenomenon. Two concentrations of NaCl (0.5 g/kg and 1.5 g/kg of soil) were employed in this study. Results revealed that the higher concentration of NaCl promoted disappearance of chloride. Other than option 9, all other options recorded significant percentage organic carbon removal at day 14. However, the removal stabilized by day 28. The inference from these results is that olive oil, the source of organic carbon was easily biodegraded in the vadose zone. The total heterotrophic bacterial count of all the options except 9 shows an increase with time. The results suggest that the population of aerobic bacteria at soil depth 6 m is sufficient to bring about over 90% organic carbon removal. Biodegradation will however be faster in surface soils where aerobic bacteria predominate and oxygen is more readily available to the resident bacterial flora. The general trend of other results indicates that the values of most components analysed declined steadily except moisture content and Total Heterotrophic Bacteria count which showed increased values. The decrease in chloride and sulphate are much smaller than the other components. A summary of the qualitative assessment of the various site specific characteristics that favour natural attenuation is shown in **Table 6**. The reduction in the values of properties in the dry season is greater than in the wet season although the seasonal differences vary within very narrow ranges. The results also show that with appropriate physicochemical properties such as the presence of biodegradable organic carbon, the availability of a nitrogen source, etc., biodegradation can take place even in deep soils (6 m deep) under limited oxygen supply. The population of resident heterotrophic bacterial flora in the subsurface soil is also much smaller than in surface soil. Microorganisms in such subsurface environments degrade organic substances slowly when compared to the surface environments [22]. At the proposed landfill site, the leachate degradation in the vadose zone and below is expected to be slow as a result of little oxygen. However, the anaerobic and microaerophilic anaerobic bacteria in the soil at this depth will facilitate the biodegrading process.

## 6. Summary and Conclusion

The results of this study may be summarized as follows:

- 1) The soils of the vadose zone (0.5 - 6.5 m) at the landfill site are mainly brownish clayey sands of low permeability. pH of the groundwater implies that it is basic in nature.
- 2) The population of aerobic bacteria within the zone is sufficient to bring about over 90% organic carbon removal.
- 3) The major constituents of the leachate such as ammonia/organic nitrogen, phosphate and organic carbon were completely degraded within 28 days while less than 50% of sulphate and chloride were degraded within the same period.
- 4) Although the decrease in the components of the leachate is generally higher in the dry season than in the wet season, the seasonal variations are quite insignificant.

**Table 6.** Qualitative assessment of the attenuation capacity of the site.

Parameter	site conditions	Values/conditions favourable for natural attenuation	Ranking (natural attenuation)
Soil Permeability	Sands; $K = 10^{-2} - 10^{-4}$ cm/sec	Moderate to high (i.e. sands, gravels.) $K > 10^{-4}$ cm/sec	Favourable
Groundwater gradient	0.4%	Low to moderate	Favourable
Depth to groundwater (thickness of vadose zone)	>8 m	Moderate to deep	Favourable
Recharge	Mean annual rainfall = 2500 m; recharge rate = 0.83 m/yr	High	Favourable
Moisture content	12% - 35%	Moderate (avoid excessively wet or dry soil)	Favourable
Dissolved oxygen (DO)	4.8 - 5.8 mg/l	Minimum is 1 - 2 mg/l	Favourable
Soil/water PH	6.97 - 7.13	6 - 8	Favourable
Total Heterotrophic (TH)	$1 \times 10^4 - 8.3 \times 10^5$	-	Favourable
Hydrocarbon Degraders (HD)	$3 \times 10^4 - 1 \times 10^5$	$>1 \times 10^5$	Favourable
Ratio of TH/HD	20% - 60%	HD is significant percentage of TH	Slightly Favourable



Results of the assessment of the various site specific characteristics reveal that both hydrogeological and chemical conditions are very favorable for natural attenuation while the *in-situ* biological conditions appear to be less favourable. The overall site conditions strongly indicate that the potential natural attenuation capacity of the vadose zone is moderate to high.

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